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► **To cite this version:**

Jean-claude Tourneur, Joël Meunier. Variations in seasonal (not mean) temperatures drive rapid adaptations to novel environments at a continent scale. *Ecology*, 2020, 101 (4), pp.e02973. 10.1002/ecy.2973 . hal-02471098

HAL Id: hal-02471098

<https://univ-tours.hal.science/hal-02471098v1>

Submitted on 20 May 2020

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1 **Variations in seasonal (not mean) temperatures drive rapid adaptations to**
2 **novel environments at a continent-scale**

3 ***Running title:*** Climate adaptation in earwigs

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12 **ABSTRACT**

13 The recent development of human societies has led to major, rapid and often inexorable
14 changes in the environment of most animal species. Over the last decades, a growing number
15 of studies formulated predictions on the modalities of animal adaptation to novel or changing
16 environments, questioning how and at what speed animals should adapt to such changes,
17 discussing the levels of risks imposed by changes in the mean and/or variance of temperatures
18 on animal performance, and exploring the underlying roles of phenotypic plasticity and genetic
19 inheritance. These fundamental predictions, however, remain poorly tested using field data.
20 Here, we tested these predictions using a unique continental-scale data set in the European
21 earwig *Forficula auricularia* L, a univoltine insect introduced in North America one century
22 ago. We conducted a common garden experiment, in which we measured 13 life-history traits
23 in 4158 field-sampled earwigs originating from 19 populations across North America. Our
24 results first demonstrate that 10 of the 13 measured life-history traits are associated with two
25 sets of variations in seasonal temperatures, i.e. winter-summer and autumn-spring. We found,
26 however, no association with the overall mean monthly temperatures of the invaded locations.
27 Furthermore, our use of a common garden setup reveals that the observed patterns of variation
28 in earwigs' life-history traits are not mere plastic responses to their current environment, but
29 are either due to their genetic background and/or to the environmental conditions they
30 experienced during early life development. Overall, these findings provide continent-scale
31 support to the claims that adaptation to thermal changes can occur quickly (in less than 100
32 generations), even in insects with long life cycles, and emphasize the importance of variation
33 in seasonal temperature over mean population temperatures in climate adaptation.

34 **Keywords:** Temperature, Adaptation, Reproductive strategy, Climate change, Invasion,
35 Dermaptera

36 INTRODUCTION

37 Adaptation to novel and changing environments is a keystone in our current understanding of
38 species distribution, abundance and evolution (Holt 1990). Over the last century, the
39 development of human activities has led to rapid and often inexorable changes in the
40 environment of many living organisms (Parmesan 2006). These changes occur, for instance,
41 due to human trade and transit, which increasingly favours the transport (invasion) of a large
42 number of animal and plant species out of their native area (Hulme 2009), and possibly bring
43 them to novel environments with unfamiliar biotic and/or abiotic properties (Jeschke and
44 Strayer 2005). These changes can also occur in the native area of these organisms due to the
45 global climate change, a phenomenon that has accelerated over the last decades and now
46 reaches an unprecedented speed (Meehl and Tebaldi 2004, Williams et al. 2007). Because the
47 extent and multiplicity of these rapid environmental changes challenge the ability of resident
48 populations to track them and adapt their life histories accordingly, understanding the nature
49 of animals' adaptation to novel and changing environment is considered a critical, ongoing and
50 challenging question in ecology (Parmesan 2006, Hill et al. 2016, 2019, Courchamp et al.
51 2017).

52 Among the multiple parameters that can vary when a species is exposed to a novel or
53 changing environment, the mean and seasonal temperatures can be particularly critical. This is
54 because the severity and duration of changes in the temperatures experienced by an individual
55 during its entire or a specific period of its life cycle can alter a great number of its life history
56 traits and ultimately modify its own fitness. For instance, extended winter durations or exposure
57 to warmer environmental conditions can shape the size and time of first reproduction (Fretwell
58 1972, Altizer et al. 2006), the duration of foetal development and levels of immune activity
59 (Körner et al. 2018, Zhang et al. 2019), as well as the mobility, morphology and metabolism
60 of individuals (Adamczewski et al. 1993, Danks 2000, Polidori et al. 2019).

61 Over the last decades, a great number of modelling and theoretical approaches have
62 been developed to better understand the nature and extent of animals' adaptation to novel
63 temperatures (Parmesan 2006). These studies formulated key predictions on how and at what
64 speed animals should adapt to such changes, on the respective importance of an increase in the
65 overall mean temperature and/or seasonality of a population on animal performance, as well as
66 on the underlying roles of phenotypic plasticity and genetic inheritance in adaptation (Nylin and
67 Gotthard 1998, Kingsolver et al. 2013, Paaajmans et al. 2013, Gilbert et al. 2014, Merilä and
68 Hendry 2014, Levis and Pfennig 2016, Williams et al. 2017, Corl et al. 2018, Fox et al. 2019,
69 Rohner et al. 2019). For instance, these studies suggest that adaptation to climate change should
70 be rapid in organisms with fast development and short life-cycles, as found in many arthropods,
71 whereas it should be slower in organisms exhibiting delayed development and long life-cycles,
72 as found in many vertebrates. Species should also be less sensitive to changes in seasonality
73 compared to changes in overall mean temperatures when they are endotherms and/or when
74 their entire life-cycle occur within a single season, whereas the opposite pattern is expected
75 when they are ectotherms and/or have a life-cycle encompassing several seasons. Finally,
76 phenotypic plasticity is often considered a keystone of rapid adaptation to environmental
77 changes, whereas fixed and inherited patterns of adaptation are often thought to secondarily
78 derive from the canalization of ancestral plastic variation (Chevin et al. 2010).

79 Although central in our current understanding of animal's adaptation to novel
80 temperatures, these fundamental predictions remain poorly tested in the field (Janion-
81 Scheepers et al. 2017, Blanckenhorn et al. 2018). This is probably because such field data are
82 difficult to collect, as it typically requires measuring variation in life-history traits across
83 multiple natural populations, over several years, and under different climates. A powerful
84 alternative consists in using field data of introduced species that recently invaded large
85 geographic areas exhibiting a broad diversity of thermal constraints (Huey et al. 2000, Bellard

86 et al. 2016). The European earwig *Forficula auricularia* Linnaeus (Dermaptera: Forficulidae)
87 is one of these species. This insect exhibits a broad native range extending across Europe, Asia
88 and northern Africa (Lamb and Wellington 1975) from which it has been introduced to
89 Australia, New Zealand, East Africa, East Indies, Tasmania and North America (Frank 1918,
90 Guillet et al. 2000, Quarrell et al. 2018, Hill et al. 2019). Its presence in North America was
91 first reported on the Pacific coast in Seattle (WA) in 1907, and then on the Atlantic coast in
92 Newport (RI) in 1911 and in Vancouver (BC) in 1919 (Crumb et al. 1941). From these
93 introductory foci (Figure 1), historical data suggests that *F. auricularia* first spread along the
94 coasts to cover areas ranging from British Columbia to California and from Newfoundland to
95 South Carolina, and then spread to the interior of the continent in both United States of America
96 and Canada (Crumb et al. 1941, Tourneur 2017)(Historical records are detailed in Appendix
97 S1: Table S1). Given that this species produces only one generation per year (Tourneur and
98 Gingras 1992, Meunier et al. 2012), these data reveal that its successful colonization of North
99 America and thus its adaptation to a broad diversity of thermal environments occurred in less
100 than 100 generations.

101 In this study, we conducted a common garden experiment to explore how *F. auricularia*
102 responded to the different thermal environments encountered during their North American
103 invasion over the last century, i.e. in less than 100 generations. Because the univoltine life cycle
104 of the European earwig encompasses all seasons and temperatures (Lamb 1976, Meunier et al.
105 2012), it has long been thought that annual mean temperatures and/or seasonality could be
106 major constraints in its invasive success (Vancassel 1984, Hill et al. 2019). However, it remains
107 unclear whether this species can mitigate these thermal constraints, and whether it does so by
108 adapting its life cycle and life-history traits (Ratz et al. 2016, Tourneur 2018).

109 Here, we 1) tested whether and how thermal constraints of the invaded locations
110 selected for specific developmental, reproductive and fitness traits (later called life history

111 traits), 2) identified the thermal constraints to which they adapted and 3) investigated the role
112 of phenotypic plasticity in this adaptation. From 1988 to 1995, we field-sampled individuals
113 originating from 19 populations located from the East to the West coasts, maintained them
114 under standard laboratory conditions and measured the properties of the 1st and 2nd (generally
115 terminal) clutches produced by each female in terms of egg laying date, egg number, egg
116 development time and number of newly hatched larvae. We also recorded the reproductive
117 strategy of the females (iteroparity versus semelparity), their reproductive outcome (total
118 number of eggs and larvae produced over lifetime), as well as the experimental survival
119 duration of the field-sampled males and females. To identify which thermal constraints the
120 tested earwigs adapted to, we tested whether our measurements could be explained by the
121 results of a principal component analyses (PCA) of the mean monthly temperatures of each
122 population. This process characterizes patterns of variation among populations' temperatures
123 without *a priori* definitions of their associations, i.e. without predetermining the focus on
124 overall mean temperatures and/or variation of temperatures between seasons or months.
125 Generating these temperature predictor variables without a priori definitions of their
126 composition was important, as both temperatures' average and seasonal variability can shape
127 animal life histories (Kingsolver et al. 2013). If *F. auricularia* individuals adapted their life-
128 cycle and life-history traits to the mean temperatures and/or seasonal temperature variation of
129 the population in which they have been sampled (and if this adaptation is determined by their
130 genetic background and/or early life experience), we predict these traits to covary with the
131 overall mean temperatures and/or variation in seasonal temperatures of their population (i.e. all
132 sampled populations should show different performance in the common garden). Conversely,
133 if earwig life-history traits are independent of the thermal environment of the population in
134 which they have been sampled (i.e. no adaptation) and/or are plastic to their current thermal
135 environment, we predict no apparent association between the traits measured in our field-

136 sampled individuals and the seasonal temperature variation of their populations (i.e. all
137 sampled populations should show similar performance in the common garden).

138 **MATERIAL AND METHODS**

139 **Earwig biology**

140 The life cycle of the European earwig generally starts with the emergence of new adults in late
141 spring to early July (with variation among populations). These adults form groups of up to
142 several hundred individuals, in which both males and females typically mate with several
143 partners (Weiß et al. 2014, Sandrin et al. 2015, Tourneur 2017). Females then burrow in the
144 ground from mid fall to early winter and build a nest where they lay their first clutch of eggs.
145 After egg laying, females stop their foraging activity and provide extensive forms of egg care
146 until hatching (Gingras and Tourneur 2001, Boos et al. 2014, Koch and Meunier 2014, Thesing
147 et al. 2015, Diehl and Meunier 2018, Körner et al. 2018). The eggs of this first clutch hatch in
148 spring and mothers remain with their newly hatched larvae for several weeks, during which
149 mothers provide larvae with multiple forms of care (Gingras and Tourneur 2001, Kölliker et
150 al. 2015, Kramer et al. 2015) and larvae exhibit forms of sibling cooperation (Falk et al. 2014,
151 Kramer et al. 2015, Kramer and Meunier 2016, Körner et al. 2016). A few weeks later, the
152 family unit is naturally disrupted. While larvae continue their development to adults in new
153 social groups, some females produce a second clutch of eggs (i.e. iteroparous as compared to
154 semelparous females), which will also receive pre- and post-hatching care and will hatch in
155 late spring (Lamb and Wellington 1975, Meunier et al. 2012, Ratz et al. 2016). All females
156 generally die during the following summer (Albouy and Caussanel 1990).

157 Whereas the European earwig is often considered a pest control in its native range
158 (Moerkens et al. 2012), it has been described as an agricultural pest and a nuisance of human

159 habitations in its newly colonized area (Crumb et al. 1941, Lamb and Wellington 1975, Walker
160 et al. 1993, Quarrell et al. 2016, 2018). Moreover, this species has been suggested to have partly
161 drove the decline of threatened and endangered invertebrates in America, such as the El
162 Segundo Blue Butterfly *Euphilotes bernardino allyni* and the Valley Elder-berry Longhorn
163 Beetle *Desmocerus californicus dimorphus* (Quarrell et al. 2018).

164 **Earwig sampling and laboratory rearing**

165 All *F. auricularia* individuals were collected over 7 years among 19 natural populations located
166 across North America (Figure 1, Table 1). These individuals were mostly collected as adults
167 using wooden traps (Tourneur 2018) between July and August, and were immediately setup in
168 glass containers (Mason Jars Company, Erie, Pennsylvania, United States of America) in
169 groups of 20 to 30 individuals. These containers received two sheets of creased toilet paper as
170 resting places for earwigs, and were then transported to the laboratory in Montreal, Canada.
171 Upon their arrival, containers were deposited in a shelf covered by a shelter and maintained
172 under the natural outdoor conditions of Montreal. During their transport and outdoor
173 maintenance, containers received an *ad libitum* amount of carrots and pollen as a food source
174 for earwigs, and were supplied with water by means of a cotton pad regularly soaked in water.
175 This setup allowed earwigs to perform non-controlled mating and to live in groups, which is
176 similar to their natural living conditions (Weiß et al. 2014, Sandrin et al. 2015, Kohlmeier et
177 al. 2016, Körner et al. 2018).

178 One to two months later (between the 7th and the 19th day of October of each year), we
179 used 4158 of these field-sampled individuals to set up 2079 mating pairs (from 17 to 356 pairs
180 per population, see Table 1), in which we subsequently measured 13 life-history traits (see
181 below). These pairs were set up in Petri dishes (diameter 10 cm) lined with a thin layer of moist
182 sand, and in which food was changed and substrate humidified once a week. Each Petri dish
183 was then transferred in a climate chamber and then maintained at 10 ± 1 °C, a temperature

184 close to the overall median temperature of the 19 sampled populations (i.e. 9.5°C, see Appendix
185 S1: Table S2). Food was removed at egg laying to mimic the natural end of earwigs' foraging
186 activity (Kölliker 2007). At egg hatching, we discarded all newly emerged larvae from the
187 experiments to trigger a novel ovarian cycle in the mothers and allow their production of a
188 subsequent clutch (Vancassel and Foraste 1980, Meunier et al. 2012). We then maintained the
189 pairs under the rearing conditions described above until our experiment ended, i.e. either one
190 year after the beginning of our laboratory setup or at the death of the adult males and females.
191 Overall, 3927 of the 4158 (94.4%) tested individuals died within the year following the
192 beginning of our experiments, a value in line with previous data on *F. auricularia* lifespan
193 (Albouy and Caussanel 1990). Note that recent studies revealed that North American *F.*
194 *auricularia* encompasses two genetic subspecies with no apparent mixing of their populations
195 (Wirth et al. 1998, Quarrell et al. 2018, Tourneur 2018). Although these subspecies were not
196 considered in our analyses (our data were collected before the publication of these genetic
197 analyses), the continuous distribution (unimodal data) of the life history traits measured across
198 populations (Figures 2 to 4) suggests an absence of subspecies-specific values regarding these
199 measurements. The potential co-occurrence of the two subspecies in our data set is thus
200 unlikely to bias our study and its main conclusions.

201 **Measurements of the life-history traits**

202 For each mating pair, we measured 13 life-history traits encompassing the properties of the
203 resulting 1st and 2nd clutches (when present), the reproductive strategy and reproductive
204 outcomes of each female, as well as the experimental survival duration of both field-sampled
205 males and females. These properties were obtained by recording the date of egg production,
206 counting the number of eggs produced, calculating the duration of egg development until
207 hatching (in days) and finally counting the number of larvae at egg hatching in both 1st and 2nd
208 clutches (when present). The reproductive strategies and reproductive outcomes of females

209 were obtained by recording whether they were semelparous or iteroparous (i.e. produced one
210 or two clutches in their lifetime, respectively), and by counting the total number of eggs and
211 larvae produced per female during their lifetime. Finally, we measured the experimental
212 survival duration of adults by counting the number of days each male and female survived after
213 October 1st of the year of field sampling. Although our measurement of survival duration does
214 not necessarily reflect adults longevity, as individuals could have different ages at field
215 sampling (see discussion), it nevertheless provides important insights into the period at which
216 males and females of each population die during the year. Note that 8.1% and 5.4% females
217 from Santa Cruz and Asheville, respectively, produced a third clutch. This third clutch was not
218 considered in the present study, as our experiment ended before their hatching.

219 **Extraction of mean and seasonal temperatures of each population**

220 We extracted the mean monthly temperature of the 19 studied populations using their GPS
221 coordinates (Table 1) and the Worldclim database v2.0 (<http://www.worldclim.org/>) with a
222 spatial resolution of 30 seconds. The mean temperatures provided by the Worldclim database
223 are calculated over 30 years, from 1970 to 2000. To reduce dimensionality of co-varying
224 temperatures in our data set while characterizing the variation in seasonal temperature of each
225 population without *a priori* definitions of their composition, we then conducted a Principal
226 Component Analysis (PCA) on the set of 12 mean monthly temperatures per population
227 (Appendix S1: Table S2). This analysis provided us with 12 orthogonal principal components
228 (PCs), out of which we retained the first three PCs (total variance explained = 98.6%, Table 2).
229 The first component (PC1) was positively loaded by almost all monthly temperatures, therefore
230 positively reflecting the overall mean temperature of a population. The second component
231 (PC2) revealed variation in seasonality between February on one hand, and June, July, and
232 August on the other hand. In particular, high values of PC2 reflected populations with cold
233 February (winter) and warm summer, whereas small values of PC2 reflected populations with

234 warm February (winter) and cold summer. Finally, the third component (PC3) captured
235 variation in seasonality between October and November on one hand, and April and May on
236 the other hand. High values of PC3 therefore characterized populations with cold autumn and
237 warm spring, whereas small values of PC3 characterized populations with warm autumn and
238 cold spring.

239 **Statistical analyses**

240 To test whether *F. auricularia* adapt their life-cycle and life-history traits to North American
241 temperatures, we conducted a series of 12 linear models (LM in R) and one generalized linear
242 model (GLM in R) – see Table 3. In the 12 LMs, the three selected PCs and their interactions
243 were entered as explanatory variables (PC1, PC2 and PC3), whereas the response variable was
244 either egg laying date, egg number, egg development time and larvae number for the 1st or 2nd
245 clutches (for a total of 8 LMs), the total number of eggs or larvae produced, or the survival
246 duration of males or females. Note that both egg laying date and adult survival duration were
247 calculated using October 1st as day 0. In the GLM, the response variable was the ratio of
248 iteroparous females per population, which was entered using the command *cbind* in R (to
249 weight each ratio by the sample size of its population) and fitted to a binomial error distribution
250 corrected for overdispersion. In all our statistical models, the response variables were the mean
251 values of each measured trait per population. They were also checked for homoscedasticity and
252 normality of residuals, as well as simplified stepwise by removing all non-significant
253 interaction terms (all $P > 0.05$). To correct for inflated Type-I errors due to multiple testing
254 (and provide an experiment-wide Type I error rate of 5%), all P -values were adjusted using
255 False Discovery Rate (FDR) correction (Benjamini and Hochberg 1995). All analyses were
256 conducted using the software R v3.5.1 loaded with the packages *raster*, *FactoMineR*, *rsq* and
257 *rcompanion*.

258 RESULTS

259 The 19 study sites greatly varied in their mean and seasonal variation of temperature (Appendix
260 S1: Table S2), as well as in the mean values of the 13 traits measured in their sampled
261 individuals (Figures 2 to 4; Appendix S1: Tables S3 to S5). Mean monthly temperatures overall
262 ranged from 22.9°C (July in Saluda) to -10.1°C (January in Montreal), while thermal
263 amplitudes over a year ranged from 30.7°C (Montreal) to 7.9°C (Santa Cruz). For the traits
264 measured in the 1st clutches, the mean \pm SE dates of egg production ranged from 47.8 \pm 0.8 to
265 132.6 \pm 3.3 days after the 1st of October, the mean \pm SE number of eggs per clutch from 23.2
266 \pm 1.9 to 66.0 \pm 2.3, the mean \pm SE egg development time from 42.2 \pm 1.0 to 71.4 \pm 0.8 days
267 and the mean \pm SE number of larvae per clutch from 11.6 \pm 2.0 to 44.8 \pm 3.2. For the 2nd
268 clutches, the mean \pm SE dates of egg production ranged from 142.0 to 248.2 \pm 5.3 days after
269 the 1st of October, the mean \pm SE number of eggs from 14.0 to 38.4 \pm 1.7, the mean \pm SE egg
270 development time from 10.0 \pm 3.0 to 63.7 \pm 4.6 days and the mean \pm SE number of larvae from
271 0 to 17.7 \pm 8.8. Finally, the total number of eggs \pm SE produced ranged from 28.1 \pm 2.6 to 83.4
272 \pm 2.9, the total number of larvae \pm SE produced from 13.0 \pm 2.2 to 46.3 \pm 3.5, the proportion
273 of iteroparous females from 0 to 70.8%, and the mean survival duration \pm SE of males and
274 females from 82.0 \pm 15.4 to 299.8 \pm 8.6 days and from 146.0 \pm 4.7 to 322.5 \pm 7.6 days after
275 the 1st of October, respectively.

276 Of the 13 measured traits, 10 varied together with the seasonal temperature variation of
277 the population of origin (Table 3). Five of these 10 traits were exclusively associated with PC2
278 (February-summer temperatures), two traits were exclusively associated with PC3 (autumn-
279 spring temperatures), and three traits were associated with both PC2 and PC3. By contrast, no
280 traits were associated with PC1 (overall mean temperatures). The associations with PC2
281 revealed that populations with cold February and warm summers (high PC2 values) had
282 females that produced their 1st clutch of eggs earlier and these eggs had longer development

283 time compared to populations exhibiting warm February and cold summers (low PC2 values,
284 Figure 2). Similarly, females from the former populations were less likely to produce a second
285 clutch (i.e. to be iteroparous, Figure 3) and when they did so, their 2nd clutch eggs were less
286 numerous (Figure 3) and showed longer development time (Figure 3). Moreover, females and
287 males from populations with cold February and warm summers lived fewer days compared to
288 adults from warm February and cold summers (Figure 4). On the other hand, the effects of PC3
289 reveal that populations exhibiting cold autumns and warm springs (high PC3 values) had
290 females that produced their 1st clutch of eggs later in the season and these eggs were less
291 numerous compared to females from populations with warm autumns and cold springs (low
292 PC3 values, Figure 2). Females from populations with high PC3 values also had 2nd clutch eggs
293 that exhibited a shorter developmental time (Figure 3), they produced an overall lower number
294 of eggs (Figure 4) and had males with a longer survival duration (Figure 4). By contrast, PC1,
295 PC2 and PC3 did not shape the number of 1st clutch larvae, as well as their total number and
296 the dates of 2nd clutch egg laying (Figures 2, 3 and 4; Table 3).

297 **DISCUSSION**

298 In this study, we demonstrate that the successful invasion of the European earwig across North
299 America came with multiple changes in their life-history traits, and that these changes are
300 associated with seasonal variation in the temperatures - but not with overall mean temperatures
301 - of the invaded areas. In particular, our data from 19 populations revealed an association
302 between seasonal temperature variation and females' timing of first reproduction, reproductive
303 strategy and investment into egg production, as well as between seasonal temperature variation
304 and the experimental survival duration of both males and females. By contrast, we found no
305 association between seasonal temperature variation and both the timing of second clutch
306 reproduction and the total number of larvae produced per female.

307 Our data first showed that females produced their first clutch of eggs earlier when they
308 came from populations facing warm summers and/or warm autumns (PC2 and PC3,
309 respectively), and were less likely to produce a second clutch in populations with cold
310 February. A plastic response to warm temperatures on egg laying date could be expected in
311 nature: adult earwigs typically develop and mate during summer and autumn, so that warm
312 temperatures during these seasons could accelerate their reproductive physiology (as shown in
313 other insect species, Singh et al., 2018) and thus accelerate egg laying (Tourneur 2018).
314 Similarly, cold Februaries might slow down the development of 1st clutch eggs and thus extend
315 the corresponding period of egg care. This, in turn, might inhibit females' physiological
316 transformation to produce a second clutch (Vancassel 1984, Gingras and Tourneur 2001,
317 Tourneur 2018, Körner et al. 2018). However, our results were obtained under common garden
318 conditions, which reveals that the observed effects of the variation in seasonal temperature on
319 egg laying dates are not a plastic response to their current environment, but are either due to
320 the environment experienced during their early life development (i.e. before field sampling),
321 or due to an inherited basis that possibly emerged through canalization (Nylin and Gotthard
322 1998, Van Buskrik and Steiner 2009). It has been proposed that traits tightly linked to fitness
323 are more strongly canalized due to past stabilizing selection (Falconer 1990). Our findings may
324 therefore suggest that the observed patterns in the timing of first reproduction and females'
325 reproductive strategy may have first emerged as a plastic response to the thermal constraints of
326 the different localities, then diverged between populations through canalization and ultimately
327 become inherited traits – all this in a maximum of 100 generations. Further experiments with
328 naïve individuals would be required to rule out an effect of early life experience.

329 Our data also reveal that seasonal temperature variation is associated with lifetime egg
330 production, but not with lifetime larvae production. In particular, the total number of eggs
331 produced per female decreased with decreasing autumn temperatures, whereas this association

332 vanished with larvae number. This apparent discrepancy suggests that eggs from populations
333 with the warmest autumns suffered higher mortality during development. A first explanation
334 of this phenomenon could be that these eggs are of a lower quality and/or that their mothers
335 were less efficient in egg care, a process that is essential to ensure egg development until
336 hatching in earwigs (Boos et al. 2014, Van Meyel et al. 2019). Whereas both effects should be
337 tested in future studies, previous results may suggest that a population-specific efficiency in
338 egg care is unlikely in this species, as a previous study showed that maternal investment in
339 post-hatching care is not population-specific, at least in Europe (Ratz et al. 2016). Another
340 explanation is that females consumed a larger proportion of their clutch in populations with the
341 warmest compared to the coldest autumns. Filial egg consumption is a common phenomenon
342 in insects (Elgar and Crespi 1992) and it has been recently reported in several Dermapteran
343 species, such as the species studied here (Koch and Meunier 2014, Van Meyel et al. 2019) and
344 the maritime earwig *Anisolabis maritima* Bonelli (Miller and Zink 2012). In the European
345 earwig, this phenomenon has been proposed to reflect an adaptive strategy to limit female
346 weight loss during the period of egg care (i.e. when they stop all other foraging activities) and
347 by doing so, to reallocate resources into post-hatching care and/or into a 2nd oogenesis cycle
348 (Koch and Meunier 2014, Tourneur 2018). Given that females lay eggs earlier in populations
349 with the warmest autumns, this increased egg consumption could be an adaptive strategy to
350 limit the cost of tending newly hatched offspring earlier in the season (middle of winter) when
351 food sources are scarce or absent. If this hypothesis holds true, it would suggest that filial egg
352 cannibalism could be a strategy that *F. auricularia* females have evolved to better cope with
353 warmer autumns.

354 In addition to the above findings, our results show that the survival duration of both
355 males and females was associated with the seasonal temperature variation of the population of
356 origin. In particular, females' and males' survival duration decreased together with warm

357 summers (and cold Februaries), while male's survival duration also decreased with warm
358 autumns (and cold springs). The first results may be a by-product of the effect of temperature
359 on their date of egg laying and/or egg hatching. We showed that females from populations
360 facing warm summers are the first to lay their eggs. Individuals from these populations might
361 thus have been the oldest at the date of our field sampling, therefore leading to the shortest
362 survival duration in our subsequent experiment. Surprisingly, there was a sex-specific effect of
363 spring (and autumn) temperatures on adult survival duration: males lived up to two times longer
364 in populations with warm compared to cold springs (as well as cold compared to warm
365 autumns), whereas this effect was absent in females. This finding may reflect sex-specific
366 sensitivity to high temperatures in terms of, for instance, physiology or expression of costly
367 behaviors. Whereas some physiological traits are known to be sex-specific in this species
368 (Kohlmeier et al. 2016, Vogelweith et al. 2017), further studies should explore the effects of
369 temperature on the observed differences. Notwithstanding its underlying mechanisms, the long
370 survival duration of males in warm spring populations opens scope for these males to mate
371 with females of the subsequent generation, as well as for a possible involvement of fathers into
372 larva care – a phenomenon reported in other insect species (Smiseth 2014). These two
373 processes remain unknown in the European earwig, but they could be of central importance in
374 their successful adaptation to climate change.

375 All our results are based on a common garden experiment, a method that is often
376 considered a powerful tool to disentangle the roles of phenotypic plasticity and genetic
377 background on adaptation (Franks et al. 2014, Stoks et al. 2014, Blanckenhorn et al. 2018).
378 Individuals reared under a common environment are typically expected to exhibit homogenized
379 life-history traits if adaptation is the outcome of phenotypic plasticity, whereas they should
380 exhibit population-specific traits otherwise. Our results are in line with the latter process for
381 the great majority of the measured traits (10 out of 13), therefore suggesting that the observed

382 associations between variation in seasonal temperature and life-history traits do not stem from
383 a plastic response to their current environment. Nevertheless, common garden experiments can
384 also have some limits: they do not necessarily prevent maternal and grand maternal effects,
385 they cannot preclude the possibility of genotype-by-environment interactions on the measured
386 life-history traits, and they are poorly efficient at shedding light on the multiple facets of
387 plasticity (e.g. some traits can be partially plastic, the plastic responses can vary in intensity
388 and slope, and plasticity may become apparent only after certain thresholds)(Franks et al. 2014,
389 Merilä and Hendry 2014, Stoks et al. 2014, Bodensteiner et al. 2019). Concluding on the
390 absence or limited role of plasticity in earwigs' adaptation to seasonality therefore needs further
391 empirical works. These works should, for instance, explore its pattern using lab-generations of
392 each population to control for potential maternal effects, or investigate the multiple facets of
393 adaptation under different common garden conditions (e.g. using different temperatures)
394 (Bodensteiner et al. 2019) and if an apparent plasticity emerge, they should demonstrate its
395 adaptive value.

396 To conclude, our results demonstrate that the spread of the European earwigs across
397 North America came with important changes in their life-history traits and life cycle, and that
398 these changes emerged in a maximum of 100 generations. Whereas we show that some of these
399 changes are by-products of novel thermal constraints (timing of first reproduction and female
400 iteroparity), we reveal that others are likely to reflect adaptive strategies to cope with different
401 autumn temperatures (egg production and the possibility of egg cannibalism). Overall, these
402 findings emphasize that adaptation of an insect with a relatively long life-cycle does not
403 necessarily operate in response to the overall mean temperatures of the invaded environments,
404 but in response to their seasonality and/or mean temperature at a specific time of their life-
405 cycle. Whether the reported adaptations are the product of population-differences in
406 energetic/metabolic constraints experienced by adults during their early development (Wong

407 and Kölliker 2014, English et al. 2016), and/or the product of an inherited genetic basis that
408 varies with seasonal temperature variation (Levis and Pfennig 2016; Corl et al. 2018; Fox et al.
409 2019), as well as whether these adaptations are similar across its worldwide distribution (Frank
410 1918, Guillet et al. 2000, Huey et al. 2000, Quarrell et al. 2018, Hill et al. 2019) will be
411 investigated in future studies. On a more general level, our findings emphasize that studying
412 invasive species can provide unique data sets to empirically and comprehensively test general
413 predictions on animals' responses to novel environmental conditions and climate change
414 (Gilbert et al. 2014, Merilä and Hendry 2014, Levis and Pfennig 2016, Hulme 2017, Fox et al.
415 2019, Rohner et al. 2019), and therefore call for their open access to the entire research
416 community - a timely task to which the present study contributes.

417 **ACKNOWLEDGMENTS**

418 This work is dedicated to the memory of Noelle Tourneur, for her constant support through
419 JCT career. We thank Maximilian Körner, Franck Dedeine and Sylvain Pincebourde, as well
420 as Eric Gangloff, Ben Phillips and Fabien Aubret (Peer Community in Evolutionary Biology
421 referees and recommenders, respectively) for their comments on a previous version of this
422 manuscript. Note that a former version of this manuscript has been recommended by Peer
423 Community in Evolutionary Biology. We also would like to thank Jean Gingras for his help
424 during laboratory observations and samplings, as well as Michel Vancassel and Maryvonne
425 Forasté for their help in the late 1980s. JCT designed the experiment, conducted field
426 samplings, and run the experiments. JM analysed the data and wrote the first version of the
427 manuscript. The final manuscript was commented and corrected by all authors. The authors
428 declare that they have no financial conflict of interest with the content of this article.

429 **REFERENCES**

430 Adamczewski, J. Z., R. J. Hudson, and C. C. Gates. 1993. Winter Energy-Balance and
431 Activity of Female Caribou on Coats-Island, Northwest-Territories - the Relative Importance
432 of Foraging and Body Reserves. *Canadian Journal of Zoology-Revue Canadienne De*
433 *Zoologie* 71:1221–1229.

434 Albouy, V., and C. Caussanel. 1990. *Dermaptères ou perce-oreilles*. Fédération Française des
435 Sociétés de Sciences Naturelles, Paris.

436 Altizer, S., A. Dobson, P. Hosseini, P. Hudson, M. Pascual, and P. Rohani. 2006. Seasonality
437 and the dynamics of infectious diseases. *Ecology Letters* 9:467–484.

438 Bellard, C., B. Leroy, W. Thuiller, J. F. Rysman, and F. Courchamp. 2016. Major drivers of
439 invasion risks throughout the world. *Ecosphere* 7:1–14.

440 Benjamini, Y., and Y. Hochberg. 1995. Benjamini Y, Hochberg Y. Controlling the false
441 discovery rate: a practical and powerful approach to multiple testing. *Journal of the Royal*
442 *Statistical Society B* 57:289–300.

443 Blanckenhorn, W. U., S. S. Bauerfeind, D. Berger, G. Davidowitz, C. W. Fox, F. Guillaume,
444 S. Nakamura, K. Nishimura, H. Sasaki, R. C. Stillwell, T. Tachi, and M. A. Schäfer. 2018.
445 Life history traits, but not body size, vary systematically along latitudinal gradients on three
446 continents in the widespread yellow dung fly. *Ecography* 41:2080–2091.

447 Bodensteiner, B. L., D. A. Warner, J. B. Iverson, C. L. Milne-Zelman, T. S. Mitchell, J. M.
448 Refsnider, and F. J. Janzen. 2019. Geographic variation in thermal sensitivity of early life
449 traits in a widespread reptile. *Ecology and Evolution* 9:2791–2802.

450 Boos, S., J. Meunier, S. Pichon, and M. Kölliker. 2014. Maternal care provides antifungal
451 protection to eggs in the European earwig. *Behavioral Ecology* 25:754–761.

452 Van Buskrik, J., and U. K. Steiner. 2009. The fitness costs of developmental canalization and
453 plasticity. *Journal of Evolutionary Biology* 22:852–860.

454 Chevin, L.-M., R. Lande, and G. M. Mace. 2010. Adaptation, plasticity, and extinction in a
455 changing environment: Towards a predictive theory. *PLoS Biology* 8:e1000357.

456 Corl, A., K. Bi, C. Luke, A. S. Challa, A. J. Stern, B. Sinervo, and R. Nielsen. 2018. The
457 genetic basis of adaptation following plastic changes in coloration in a novel environment.
458 *Current Biology* 28:2970-2977.e7.

459 Courchamp, F., A. Fournier, C. Bellard, C. Bertelsmeier, E. Bonnaud, J. M. Jeschke, and J.
460 C. Russell. 2017. Invasion Biology: Specific Problems and Possible Solutions. *Trends in*
461 *Ecology and Evolution* 32:13–22.

462 Crumb, S. E., P. M. Eide, and A. E. Bonn. 1941. The European earwig. Page USDA
463 Technical Bulletin. USDA Technical Bulletin 766.

464 Danks, H. V. 2000. Dehydration in dormant insects. *Journal of Insect Physiology* 46:837–
465 852.

466 Diehl, J. M., and J. Meunier. 2018. Surrounding pathogens shape maternal egg care but not
467 egg production in the European earwig. *Behavioral Ecology* 29:128–136.

468 Elgar, M. A., and B. J. Crespi. 1992. Cannibalism: ecology and evolution among diverse
469 taxa. Oxford University Press, Oxford.

470 English, S., T. W. Fawcett, A. D. Higginson, P. C. Trimmer, and T. Uller. 2016. Adaptive use
471 of information during growth can explain long-term effects of early Life experiences. *The*
472 *American Naturalist* 187:620–632.

473 Falconer, D. S. 1990. Selection in different environments: effects on environmental
474 sensitivity (reaction norm) and on mean performance. *Genetical Research* 56:57.

475 Falk, J., J. W. Y. Wong, M. Kölliker, and J. Meunier. 2014. Sibling cooperation in earwig
476 families provides insights into the early evolution of social life. *The American Naturalist*

477 183:547–557.

478 Fox, R. J., J. M. Donelson, C. Schunter, T. Ravasi, and J. D. Gaitán-Espitia. 2019. Beyond
479 buying time: the role of plasticity in phenotypic adaptation to rapid environmental change.
480 Philosophical Transactions of the Royal Society B: Biological Sciences 374:20180174.

481 Frank, L. M. 1918. Notes from tasmania. J Econ Ent 11:472–475.

482 Franks, S. J., J. J. Weber, and S. N. Aitken. 2014. Evolutionary and plastic responses to
483 climate change in terrestrial plant populations. Evolutionary Applications 7:123–139.

484 Fretwell, J. 1972. Populations in a Seasonal Environment. Page (P. U. Press, Ed.). Princeton,
485 NJ.

486 Gilbert, B., H. S. Greig, K. S. McCann, M. I. O’Connor, J. P. DeLong, C. D. G. Harley, V.
487 Savage, T. D. Tunney, and D. A. Vasseur. 2014. Increased temperature variation poses a
488 greater risk to species than climate warming. Proceedings of the Royal Society B: Biological
489 Sciences 281:20132612–20132612.

490 Gingras, J., and J.-C. Tourneur. 2001. Timing of adult mortality, oviposition, and hatching
491 during the underground phase of *Forficula auricularia* (Dermaptera: Forficulidae). Canadian
492 Entomologist 133:269–278.

493 Guillet, S., N. Josselin, and M. Vancassel. 2000. Multiple introductions of the *Forficula*
494 *auricularia* species complex (Dermaptera: Forficulidae) in eastern North America. Canadian
495 Entomologist 132:49–57.

496 Hill, M. P., M. Binns, P. A. Umina, A. A. Hoffmann, and S. Macfadyen. 2019. Climate,
497 human influence and the distribution limits of the invasive European earwig, *Forficula*
498 *auricularia* , in Australia. Pest Management Science 75:134–143.

499 Hill, M. P., S. Clusella-Trullas, J. S. Terblanche, and D. M. Richardson. 2016. Drivers,

500 impacts, mechanisms and adaptation in insect invasions. *Biological Invasions* 18:883–891.

501 Holt, R. D. 1990. The microevolutionary consequences of climate change. *Trends in Ecology*
502 *and Evolution* 5:590–596.

503 Huey, R. B., G. W. Gilchrist, M. L. Carlson, D. Berrigan, and L. Serra. 2000. Rapid evolution
504 of a geographic cline in size in an introduced fly. *Science* 287:308–309.

505 Hulme, P. E. 2009. Trade, transport and trouble: Managing invasive species pathways in an
506 era of globalization. *Journal of Applied Ecology* 46:10–18.

507 Hulme, P. E. 2017. Climate change and biological invasions: evidence, expectations, and
508 response options. *Biological Reviews* 92:1297–1313.

509 Janion-Scheepers, C., L. Phillips, C. M. Sgrò, G. A. Duffy, R. Hallas, and S. L. Chown. 2017.
510 Basal resistance enhances warming tolerance of alien over indigenous species across latitude.
511 *Proceedings of the National Academy of Sciences* 115:145–150.

512 Jeschke, J. M., and D. L. Strayer. 2005. Invasion success of vertebrates in Europe and North
513 America. *Proceedings of the National Academy of Sciences of the United States of America*
514 102:7198–7202.

515 Kingsolver, J. G., S. E. Diamond, and L. B. Buckley. 2013. Heat stress and the fitness
516 consequences of climate change for terrestrial ectotherms. *Functional Ecology* 27:1415–
517 1423.

518 Koch, L. K., and J. Meunier. 2014. Mother and offspring fitness in an insect with maternal
519 care: phenotypic trade-offs between egg number, egg mass and egg care. *BMC evolutionary*
520 *biology* 14:125.

521 Kohlmeier, P., K. Holländer, and J. Meunier. 2016. Survival after pathogen exposure in
522 group-living insects: don't forget the stress of social isolation! *Journal of Evolutionary*

523 Biology 29:1867–1872.

524 Kölliker, M. 2007. Benefits and costs of earwig (*Forficula auricularia*) family life.

525 Behavioral Ecology and Sociobiology 61:1489–1497.

526 Kölliker, M., S. Boos, J. W. Y. Wong, L. Röllin, D. Stucki, S. Raveh, M. Wu, and J.

527 Meunier. 2015. Parent-offspring conflict and the genetic trade-offs shaping parental

528 investment. Nature communications 6:6850.

529 Körner, M., J. M. Diehl, and J. Meunier. 2016. Growing up with feces: benefits of allo-

530 coprophagy in families of the European earwig. Behavioral Ecology 27:1775–1781.

531 Körner, M., S. Foitzik, and J. Meunier. 2018. Extended winters entail long-term costs for

532 insect offspring reared in an overwinter burrow. Journal of Thermal Biology 74:116–122.

533 Kramer, J., and J. Meunier. 2016. Maternal condition determines offspring behavior toward

534 family members in the European earwig. Behavioral Ecology 27:494–500.

535 Kramer, J., J. Thesing, and J. Meunier. 2015. Negative association between parental care and

536 sibling cooperation in earwigs: a new perspective on the early evolution of family life?

537 Journal of Evolutionary Biology 28:1299–1308.

538 Lamb, R. J. 1976. Parental behavior in the dermaptera with special reference to *Forficula*

539 *auricularia* (Dermaptera: Forficulidae). Canadian Journal of Entomology 108:609–619.

540 Lamb, R. J., and W. G. Wellington. 1975. Life history and population characteristics of the

541 european earwig, *Forficula auricularia* (Dermaptera: forficulidae), at Vancouver, British

542 columbia. Canadian Journal of Entomology 107:819–824.

543 Levis, N. A., and D. W. Pfennig. 2016. Evaluating “Plasticity-First” Evolution in Nature:

544 Key Criteria and Empirical Approaches. Trends in Ecology and Evolution 31:563–574.

545 Meehl, G. A., and C. Tebaldi. 2004. More Intense, More Frequent, and Longer Lasting Heat

546 Waves in the 21st Century. *Science* 305:994–997.

547 Merilä, J., and A. P. Hendry. 2014. Climate change, adaptation, and phenotypic plasticity:
548 The problem and the evidence. *Evolutionary Applications* 7:1–14.

549 Meunier, J., J. W. Y. Wong, Y. Gómez, S. Kuttler, L. Röllin, D. Stucki, and M. Kölliker.
550 2012. One clutch or two clutches? Fitness correlates of coexisting alternative female life-
551 histories in the European earwig. *Evolutionary Ecology* 26:669–682.

552 Van Meyel, S., S. Devers, and J. Meunier. 2019. Love them all: mothers provide care to
553 foreign eggs in the European earwig *Forficula auricularia*. *Behavioral Ecology*.

554 Miller, J. S., and A. G. Zink. 2012. Parental care trade-offs and the role of filial cannibalism
555 in the maritime earwig, *Anisolabis maritima*. *Animal Behaviour* 83:1387–1394.

556 Moerkens, R., H. Leirs, G. Peusens, T. Beliën, and B. Gobin. 2012. Natural and human
557 causes of earwig mortality during winter: Temperature, parasitoids and soil tillage. *Journal of*
558 *Applied Entomology* 136:490–500.

559 Nylin, S., and K. Gotthard. 1998. Plasticity in life-history traits. *Annual review of*
560 *entomology* 43:63–83.

561 Paaijmans, K. P., R. L. Heinig, R. A. Seliga, J. I. Blanford, S. Blanford, C. C. Murdock, and
562 M. B. Thomas. 2013. Temperature variation makes ectotherms more sensitive to climate
563 change. *Global Change Biology* 19:2373–2380.

564 Parmesan, C. 2006. Ecological and evolutionary responses to recent climate change. *Annual*
565 *Review of Ecology, Evolution, and Systematics* 37:637–669.

566 Polidori, C., C. Gutiérrez-Cánovas, E. Sánchez, J. Tormos, L. Castro, and D. Sánchez-
567 Fernández. 2019. Climate change-driven body size shrinking in a social wasp. *Ecological*
568 *Entomology*:een.12781.

569 Quarrell, S. R., J. Arabi, A. Suwalski, M. Veuille, T. Wirth, and G. R. Allen. 2018. The
570 invasion biology of the invasive earwig, *Forficula auricularia* in Australasian ecosystems.
571 *Biological Invasions* 20:1553–1565.

572 Quarrell, S. R., N. W. Davies, P. W. Walker, R. Corkrey, J. A. Smith, and G. R. Allen. 2016.
573 Identification of the putative aggregation pheromone components emitted by the European
574 earwig, *Forficula auricularia*. *Chemoecology*.

575 Ratz, T., J. Kramer, M. Veuille, and J. Meunier. 2016. The population determines whether
576 and how life-history traits vary between reproductive events in an insect with maternal care.
577 *Oecologia* 182:443–452.

578 Rohner, P. T., J. Roy, M. A. Schäfer, W. U. Blanckenhorn, and D. Berger. 2019. Does
579 thermal plasticity align with local adaptation? An interspecific comparison of wing
580 morphology in sepsid flies. *Journal of Evolutionary Biology* 32:463–475.

581 Sandrin, L., J. Meunier, S. Raveh, J.-C. Walser, and M. Kölliker. 2015. Multiple paternity
582 and mating group size in the European earwig, *Forficula auricularia*. *Ecological Entomology*
583 40:159–166.

584 Singh, S., G. Mishra, and Omkar. 2018. Plasticity in reproductive output and development in
585 response to thermal variation in ladybird beetle, *Menochilus sexmaculatus*. *Journal of*
586 *Thermal Biology* 71:180–188.

587 Smiseth, P. T. 2014. Parental care. Pages 221–242 in D. M. Shuker and L. W. Simmons,
588 editors. *The evolution of insect mating systems*. Oxford University Press, Oxford.

589 Stoks, R., A. N. Geerts, and L. De Meester. 2014. Evolutionary and plastic responses of
590 freshwater invertebrates to climate change: Realized patterns and future potential.
591 *Evolutionary Applications* 7:42–55.

592 Thesing, J., J. Kramer, L. K. Koch, and J. Meunier. 2015. Short-term benefits, but

593 transgenerational costs of maternal loss in an insect with facultative maternal care.
594 Proceedings of the Royal Society B: Biological Sciences 282:20151617.

595 Tourneur, J.-C. 2017. Epigeal phase of the biological cycle of *Forficula auricularia* Linnaeus
596 (Dermaptera: Forficulidae) in eastern Canada. Canadian Entomologist 149:600–606.

597 Tourneur, J.-C. 2018. Factors affecting the egg-laying pattern of *Forficula auricularia*
598 (Dermaptera: Forficulidae) in three climatologically different zones of North America.
599 Canadian Entomologist 150:511–519.

600 Tourneur, J.-C., and J. Gingras. 1992. Egg laying in a northeastern north american (Montréal,
601 Québec) population of *Forficula auricularia* L. (Dermaptera: Forficulidae). Canadian
602 Entomologist 124:1055–1061.

603 Vancassel, M. 1984. Plasticity and Adaptive Radiation of Dermapteran Parental Behavior:
604 Results and Perspectives. Advances in the Study of Behavior 14:51–80.

605 Vancassel, M., and M. Foraste. 1980. Le comportement parental des Dermaptères. Reprod.
606 Nutr. Dévelop. 20:759–770.

607 Vogelweith, F., S. Foitzik, and J. Meunier. 2017. Age, sex, mating status, but not social
608 isolation interact to shape basal immunity in a group-living insect. Journal of Insect
609 Physiology 103:64–70.

610 Walker, K. A., T. H. Jones, and R. D. Fell. 1993. Pheromonal basis of aggregation in
611 european earwig, *Forficula auricularia* L. (Dermaptera: Forficulidae). Journal of Chemical
612 Ecology 19:2029–2038.

613 Weiß, C., J. Kramer, K. Holländer, and J. Meunier. 2014. Influences of relatedness, food
614 deprivation, and sex on adult behaviors in the group-living insect *Forficula auricularia*.
615 Ethology 120:923–932.

616 Williams, C. M., G. J. Ragland, G. Betini, L. B. Buckley, Z. A. Cheviron, K. Donohue, J.
617 Hereford, M. M. Humphries, S. Lisovski, K. E. Marshall, P. S. Schmidt, K. S. Sheldon, Ø.
618 Varpe, and M. E. Visser. 2017. Understanding evolutionary impacts of seasonality: An
619 introduction to the symposium. *Integrative and Comparative Biology*:1–13.

620 Williams, J. W., S. T. Jackson, and J. E. Kutzbach. 2007. Projected distributions of novel and
621 disappearing climates by 2100 AD. *Proceedings of the National Academy of Sciences*
622 104:5738–5742.

623 Wirth, T., R. Le Guellec, M. Vancassel, and M. Veuille. 1998. Molecular and reproductive
624 characterization of sibling species in the european earwig (*Forficula auricularia*). *Evolution*
625 52:260.

626 Wong, J. W. Y., and M. Kölliker. 2014. Effects of food restriction across stages of juvenile
627 and early adult development on body weight, survival and adult life history. *Journal of*
628 *evolutionary biology* 27:2420–30.

629 Zhang, Y.-B., G.-F. Zhang, W.-X. Liu, and F.-H. Wan. 2019. Continuous heat waves change
630 the life history of a host-feeding parasitoid. *Biological Control* 135:57–65.

631

632 **DATA AVAILABILITY**

633 The complete data set and R script are archived in the open data repository Zenodo:
634 <https://doi.org/10.5281/zenodo.2652192>.

635

636 **Table 1 – Details of the 19 sampled populations.** The table shows the name and location of
637 each population, their GPS coordinates (Latitude, Longitude), samplings years, total number
638 of mating pair setup across years (N. pairs), and seasonal temperature variation (defined as
639 PC1, PC2 and PC3).

Populations	Country	State (USA)/Province (CDN)	Latitude	Longitude	Samplings	N. pairs	PC1	PC2	PC3
Asheville	USA	North Carolina	35.612	-82.566	1994-95	80	4.60	0.73	0.74
Charlestown	USA	Rhode Island	41.383	-71.642	1990	42	1.37	0.73	-0.84
Deschutes	USA	Oregon	44.157	-121.256	1990	17	-1.54	-2.32	-0.27
Enderby	CDN	British Columbia	50.551	-119.14	1989-90	121	-1.68	-0.02	1.14
Ennis lake	USA	Montana	45.447	-111.695	1990	36	-2.86	-0.36	0.02
Kimberley	CDN	British Columbia	49.635	-115.998	1990	94	-5.27	-0.95	0.73
Kingston	USA	Rhode Island	41.486	-71.531	1991	137	1.00	0.76	-0.75
Montreal	CDN	Quebec	45.542	-73.893	1988,1990- 95	356	-2.78	2.32	-0.07
Pointe Pelée	CDN	Ontario	41.963	-82.518	1992	47	1.25	2.31	-0.37
Revelstoke	CDN	British Columbia	50.998	-118.196	1989-90	100	-2.69	-0.11	0.95
Rocky knob	USA	Virginia	36.832	-80.345	1993-94	304	1.44	-0.07	0.48
Saluda	USA	North Carolina	35.198	-82.353	1993-95	117	5.03	0.69	0.77
Santa Cruz*	USA	California	36.926	-121.845	1991	130	5.04	-4.50	-0.57
Selkirk	CDN	Ontario	42.834	-79.932	1992-94	233	-0.69	1.35	-0.67
Selinsgrove	USA	Pennsylvania	40.832	-76.872	1993-94	134	1.76	1.83	0.27
Truro	CDN	Nova Scotia	45.372	-63.264	1988	39	-3.46	-0.13	-1.42
Vancouver	CDN	British Columbia	49.252	-123.24	1989,1991	84	0.23	-2.89	-0.05
Waterrock knob	USA	North Carolina	35.464	-83.138	1991-94	167	-1.88	-1.69	0.22
Wheatley	CDN	Ontario	42.094	-82.445	1992	52	1.13	2.31	-0.31

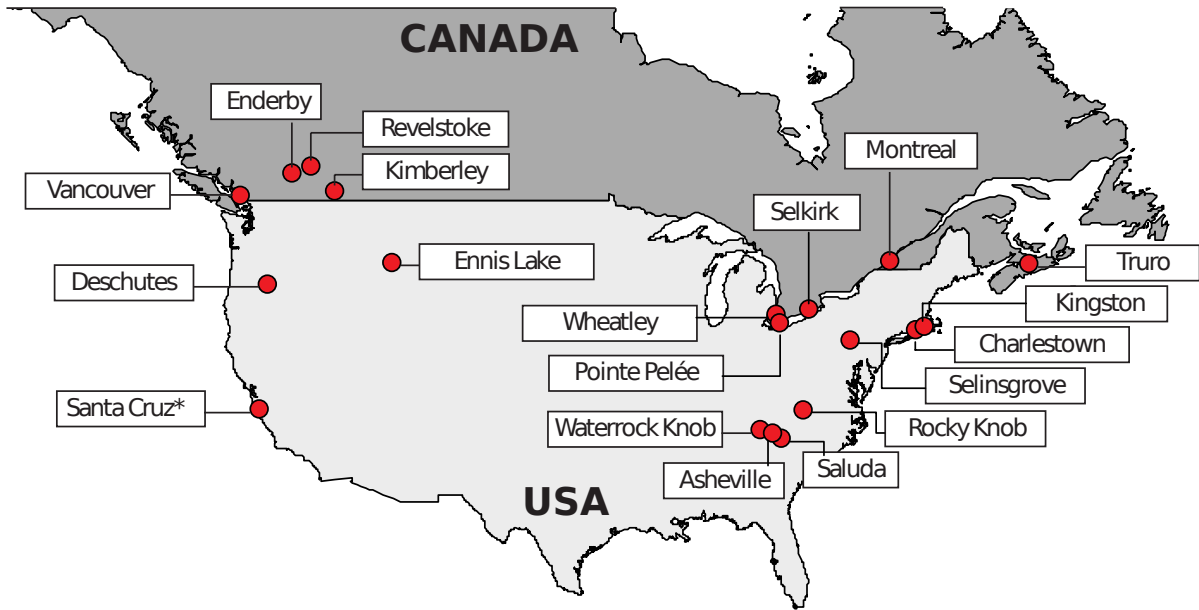
* This population was called San Francisco in (Tourneur 2018).

640 **Table 2 – Loadings of the four first principal components (PCs) reflecting combinations**
 641 **of the 12 mean monthly temperatures across populations.** The traits having significant
 642 loadings on each PC are in bold.

	PC1	PC2	PC3	PC4
Jan	0.800	-0.589	-0.066	0.083
Feb	0.716	-0.668	0.131	0.139
Mar	0.844	-0.486	0.216	0.048
Apr	0.949	-0.140	0.267	-0.082
May	0.890	0.321	0.286	-0.145
Jun	0.731	0.665	0.123	-0.060
Jul	0.547	0.823	-0.006	0.143
Aug	0.641	0.746	-0.013	0.175
Sep	0.905	0.380	-0.175	-0.019
Oct	0.951	0.019	-0.292	-0.064
Nov	0.931	-0.174	-0.296	-0.112
Dec	0.872	-0.469	-0.113	0.041
Eigenvalues	8.153	3.224	0.453	0.130
Variance explained (%)	67.9	26.9	3.8	1.1
Cumulative variance explained (%)	67.9	94.8	98.6	99.7

643 **Table 3 –Results of the statistical models on the 13 measured life-history traits.** PC1 positively reflects the overall mean temperature of a
644 population. High values of PC2 reflect populations with cold February (winter) and warm summer, and vice-versa. High values of PC3 reflect
645 populations with warm spring and cold autumn, and vice-versa. P-values significant after FDR correction (adj-P) are in bold. Note that FDR
646 correction transforms each P-value in function of its rank of statistical significance in the data set, which can lead to similar corrected p-values.
647 Model estimates (estim).

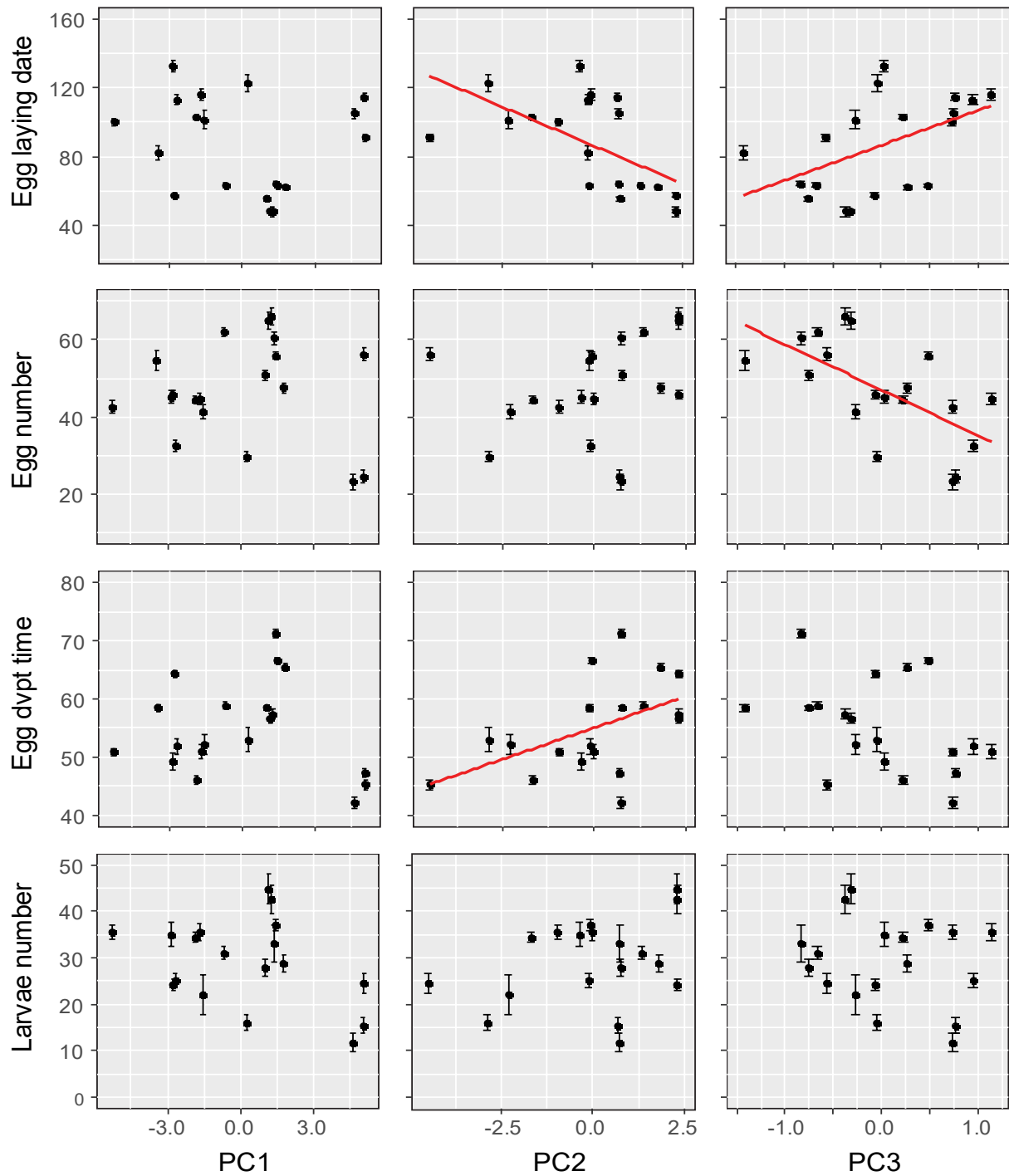
	PC1				PC2				PC3			
	estim.	SE	P	adj-P	estim.	SE	P	adj-P	estim.	SE	P	adj-P
First clutch												
Egg laying date	-1.55	1.47	0.307	0.665	-8.92	2.34	0.002	0.011	20.37	6.23	0.005	0.014
Egg number	-0.32	0.82	0.705	0.896	1.87	1.30	0.171	0.234	-11.77	3.47	0.004	0.014
Egg development time	-0.31	0.53	0.575	0.896	2.16	0.85	0.022	0.041	-4.57	2.26	0.061	0.132
Larvae number	-1.28	0.72	0.098	0.425	1.53	1.08	0.180	0.234	-3.85	3.39	0.275	0.357
Second clutch												
Egg laying date	0.79	2.01	0.700	0.896	-3.28	3.26	0.331	0.391	13.13	8.46	0.143	0.233
Egg number	0.08	0.44	0.855	0.896	-1.97	0.72	0.016	0.041	0.27	1.86	0.887	0.887
Egg development time	1.92	0.92	0.059	0.381	3.93	1.48	0.021	0.041	-19.02	4.04	0.001	0.005
Larvae number	-0.05	0.38	0.896	0.896	-1.70	0.59	0.012	0.041	2.96	1.78	0.121	0.224
General												
Total egg number	0.17	1.01	0.866	0.896	-0.15	1.60	0.929	0.929	-13.93	4.26	0.005	0.014
Total larvae number	-1.04	0.77	0.198	0.642	0.60	1.16	0.615	0.666	-3.45	3.61	0.355	0.419
Ratio of iteroparous females	0.01	0.09	0.896	0.896	-0.35	0.13	0.022	0.041	-0.48	0.40	0.254	0.357
Male longevity	-10.26	3.63	0.013	0.165	-23.93	5.77	0.001	0.011	65.49	15.39	0.001	0.005
Female longevity	-4.30	3.74	0.268	0.665	-14.80	5.94	0.025	0.041	13.63	15.86	0.404	0.437



648

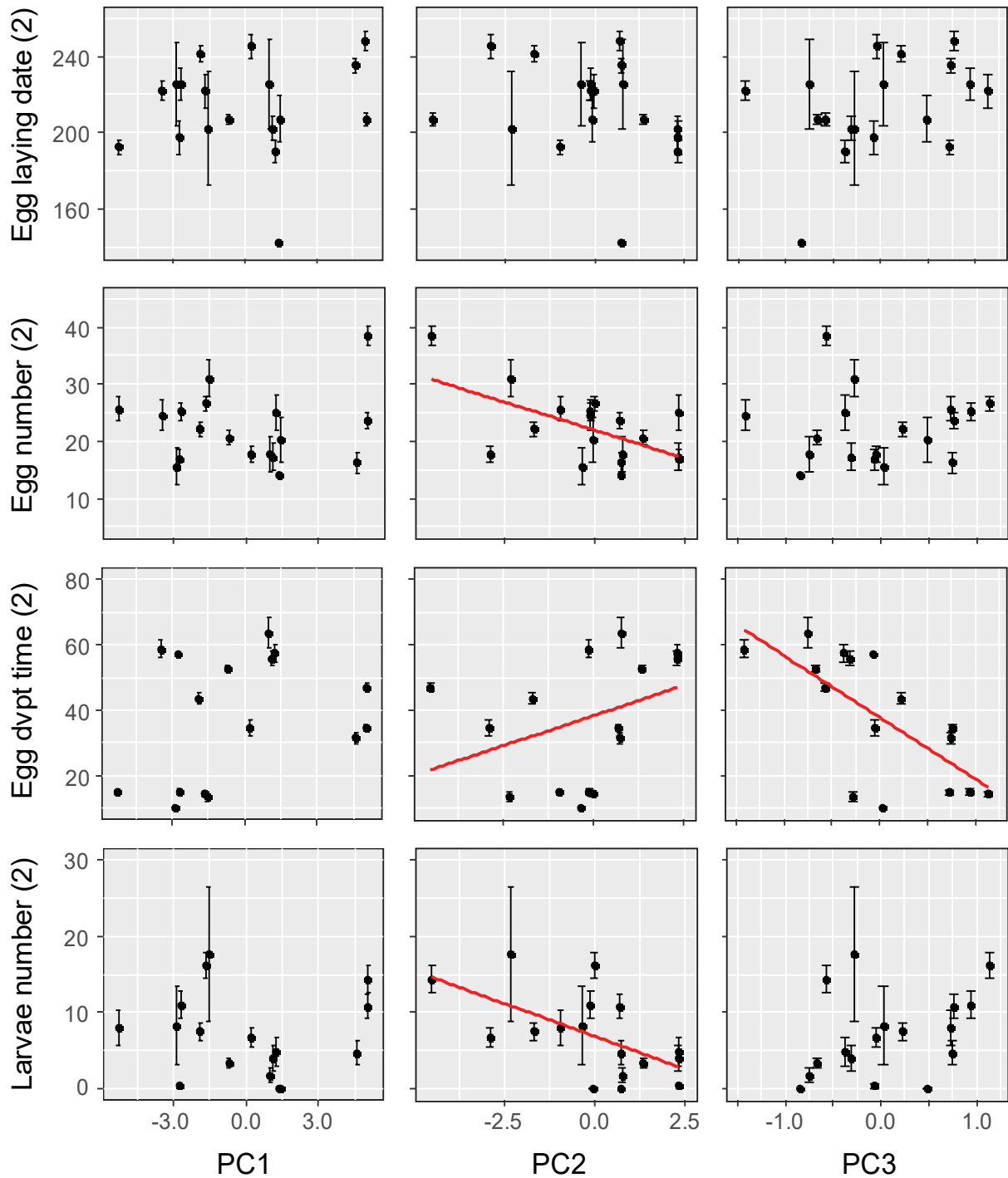
649 **Figure 1** – Map showing the 19 sampled populations across Canada (CND) and United States
 650 of America (USA). * This population was called San Francisco in Tourneur (2018).

651



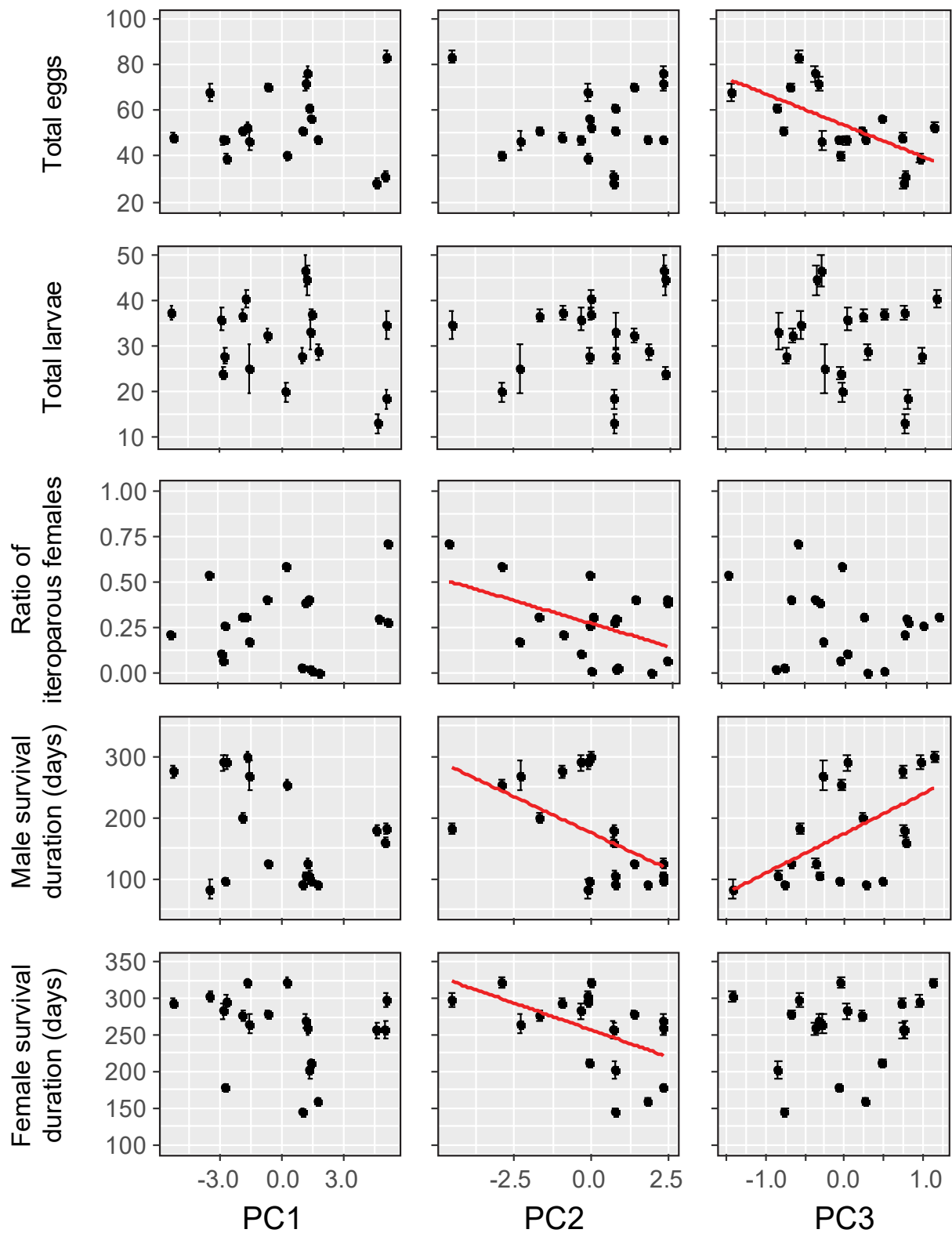
652 **Figure 2** – Associations between variation in seasonal temperatures (PC1, PC2, PC3) of the
 653 19 populations of origin and 1st clutch parameters. Red lines represent correlations significant
 654 after FDR correction. Mean values \pm SE. Egg laying date was calculated using October 1st as a
 655 reference (i.e. as day 0).

656



657

658 **Figure 3** – Associations between variation in seasonal temperatures (PC1, PC2, PC3) of the
 659 19 populations of origin and 2nd clutch parameters (when produced). Red lines represent
 660 correlations significant after FDR correction. Mean values \pm SE. Egg laying date was
 661 calculated using October 1st as a reference (i.e. as day 0).



662

663 **Figure 4** – Associations between variation in seasonal temperatures (PC1, PC2, PC3) of the

664 19 populations of origin and females' reproductive strategies and outcomes, as well as adult's

665 survival duration. Red lines represent correlations significant after FDR correction. Mean

666 values \pm SE.